

**UNIVERSIDADE FEDERAL DE SANTA MARIA
CENTRO DE CIÊNCIAS RURAIS
PROGRAMA DE PÓS GRADUAÇÃO EM CIÊNCIA DO SOLO**

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**Bioindicadores de qualidade do solo em um sistema integrado de
produção agropecuária**

**Santa Maria, RS
2016**

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**BIOINDICADORES DE QUALIDADE DO SOLO EM UM SISTEMA INTEGRADO
DE PRODUÇÃO AGROPECUÁRIA**

Tese apresentada ao Curso de Pós-Graduação em Ciência do Solo, da Universidade Federal de Santa Maria (UFMS, RS), como requisito parcial para obtenção do título de **Doutor em Ciência do Solo**.

Orientador: Prof. Dr. Rodrigo Josemar Seminoti Jacques

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Ficha catalográfica elaborada através do Programa de Geração Automática da Biblioteca Central da UFSM, com os dados fornecidos pelo(a) autor(a).

Hübert Neufeld, Ângela Denise
Bioindicadores de qualidade do solo em um sistema
integrado de produção agropecuária / Ângela Denise Hübert
Neufeld.- 2016.
74 p.; 30 cm

Orientador: Rodrigo Josemar Seminoti Jacques
Tese (doutorado) - Universidade Federal de Santa
Maria, Centro de Ciências Rurais, Programa de Pós-
Graduação em Ciência do Solo, RS, 2016

1. Bioindicadores de qualidade do solo 2. Microbiota
do solo 3. Invertebrados edáficos I. Seminoti Jacques,
Rodrigo Josemar II. Título.

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Santa Maria, RS
2016

DEDICATÓRIA

Aos meus pais Eduardo e Elvine e ao meu esposo Renan, dedico.

AGRADECIMENTOS

À Universidade Federal de Santa Maria por minha formação profissional.

Aos órgãos de fomento CNPq, CAPES, FAPERGS e Agrisus pelo financiamento das atividades de pesquisa e concessão de bolsas de auxílio.

Ao Programa de Pós Graduação em Ciência do Solo e aos funcionários Heverton, Antônio e Eunice pela ajuda nestes anos.

Ao professor Rodrigo Josemar Seminoti Jacques pela orientação durante este processo de doutoramento.

Aos professores Ibanor Anghinoni e Paulo César Faccio de Carvalho e ao grupo de pesquisa GSIPA da UFRGS pela oportunidade de realizar as avaliações na área experimental em São Miguel das Missões/RS.

À professora Zaida Antoniulli e aos meus colegas do grupo de pesquisa em biologia do solo, pelo apoio dado nas coletas e análises das amostras, em especial à Daiane e ao Willian.

Ao professor Dilmar Baretta e aos alunos da UDESC pelo apoio nas coletas.

À banca examinadora pela disponibilidade e contribuições.

Às amigas cultivadas ao longo de minha vida, que somam momentos tão importantes nesta jornada: Juliane, Renata, Daiana, Rosângela, Ana Paula e Afnan.

Aos colegas de trabalho da URI-Santo Ângelo pelo apoio.

Agradeço em especial à minha família, meus amados pais Eduardo e Elvine, meus irmãos Luciano e Wágner e minha cunhada Clara pelo carinho incondicional, apoio e compreensão em mais uma etapa tão importante de minha vida. Sem o incentivo de vocês não teria chegado até aqui, muito obrigada por tudo!

Ao meu esposo Renan, meu muito obrigada. Teu apoio, incentivo, ajuda, amparo, força e compreensão durante este período foram essenciais para esta caminhada. Te amo!

Aos meus sogros, Deli e Roque, aos cunhados Frederico, Patrícia, Filipe e Aline, e às pequenas sobrinhas, Lara e Alice, obrigada pela compreensão nos nossos momentos de distância, e pelo incentivo durante esse período.

À Deus minha eterna gratidão, por colocar em meu caminho cada um mencionado acima e pela alegria de despertar a cada dia novas curiosidades, dar novos ensinamentos e mostrar que sempre temos mais a aprender e evoluir.

"Tivesse a noite límpida milhões de estrelas,
mas equidistantes e de igual brilho, como se
admirariam o Cruzeiro do Sul, a Estrela
d'Alva, as Três Marias? A Diversidade é o
encanto da Vida."

(Leopoldo Magno Coutinho)

RESUMO

BIOINDICADORES DE QUALIDADE DO SOLO EM UM SISTEMA INTEGRADO DE PRODUÇÃO AGROPECUÁRIA

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Os sistemas integrados de produção agropecuária (SIPAs) podem melhorar os atributos químicos, físicos e biológicos do solo. Porém, a intensificação do pastejo pode suplantiar estes benefícios e causar graves prejuízos ao SIPA. A biota do solo é considerada uma boa bioindicadora ambiental e pode atestar sobre a qualidade do manejo deste sistema. O objetivo do estudo foi conhecer as implicações da intensificação do pastejo sobre a biota do solo em um SIPA de longa duração, verificar quais as condições edáficas mais influenciam os atributos biológicos, e avaliar se estes atributos podem ser utilizados como indicadores de qualidade ambiental do SIPA. O experimento é conduzido há 15 anos em uma área de 23 hectares com a sucessão *Glycine max* para produção de grãos no verão e *Avena strigosa* + *Lolium multiflorum* para o pastejo contínuo dos bovinos no inverno. Os tratamentos são constituídos pelas alturas de pastejo de 10, 20, 30 e 40 cm e por testemunhas sem pastejo. A amostragem foi realizada em quatro épocas entre 2014 a 2016, duas após o final da estação de pastejo e duas após a colheita da soja. Foram avaliados a respiração basal do solo, o conteúdo de carbono na biomassa microbiana (CBM), o quociente metabólico e a diversidade e abundância da meso e macrofauna epiedáfica, além das variáveis químicas, físicas e de cobertura vegetal para fins de correlação. No tratamento sem pastejo e nas maiores alturas da pastagem quantificaram-se a maior respiração microbiana, conteúdo de CBM, abundância e diversidade da fauna epiedáfica. Na menor altura de pastejo houve redução da cobertura vegetal e prejuízos aos atributos físicos do solo, o que resultou em menor teor de umidade e redução da atividade, abundância e diversidade dos organismos do solo. Estes resultados reforçam a necessidade de um manejo adequado da carga animal na pastagem para que não ocorra comprometimento à sustentabilidade dos SIPAs.

Palavras-chave: Microrganismo. Invertebrado do solo. Fauna edáfica. Bioindicador. Integração lavoura-pecuária. Sustentabilidade.

ABSTRACT

SOIL QUALITY BIOINDICATORS IN AN INTEGRATED CROP-LIVESTOCK SYSTEM

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Integrated crop-livestock systems (ICLS) can improve the chemical, physical and biological soil attributes. However, grazing intensification may outweigh these benefits and cause severe losses to the system. Soil biota is considered a good environmental bioindicator and can attest to the quality of the system management. This study aimed to know the implications of grazing intensification on soil biota in a long-term ICLS, to verify which soil conditions most influence biological attributes, and to evaluate if these attributes can be used as environmental quality indicators in ICLS. The experiment was started in 2001, on a 23 hectare area, with *Glycine max* in summer and *Avena strigosa*+*Lolium multiflorum* for continuous cattle grazing in winter. The treatments consisted of four sward heights (10, 20, 30, and 40 cm), plus an ungrazed area, as the control. Sampling was performed in four seasons between 2014-2016, two after the grazing season and two after soybean harvest. Soil basal respiration, microbial biomass carbon (MBC), microbial quotient, diversity and abundance of meso and macrofauna, as well as chemical, physical and vegetation cover variables were evaluated for correlation. The highest microbial respiration, MBC content, abundance and diversity of the soil fauna were quantified under moderate (20 and 30 cm) and light (40 cm) grazing intensities and under no grazing. At the high grazing intensity was a vegetal cover reduction and decreasing on the physical soil quality, which resulted in a lower soil moisture and activity, abundance and diversity reduction of soil organisms community. These results increase the need for an adequate management of pasture areas, without harming the ICLS sustainability.

Keywords: Microorganism. Soil invertebrate. Soil fauna. Bioindicator. Sustainability. Pasture areas.

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1 INTRODUÇÃO GERAL

1.1 SISTEMAS INTEGRADOS DE PRODUÇÃO AGROPECUÁRIA

Durante o último século a população mundial aumentou quatro vezes, fato possível devido à melhoria das técnicas agrícolas e industriais (TURMEL et al., 2015). Este crescimento populacional contínuo demanda um constante aumento da produção agropecuária, com manutenção da sustentabilidade dos sistemas agrícolas para as próximas gerações. A solução para essa questão está associada ao manejo sustentável do solo (LAL, 2007), com a intensificação do uso de áreas já utilizadas para a agricultura, através de alternativas que alcancem a sustentabilidade e aumentem a eficiência agropecuária (VILELA et al., 2008).

Neste sentido, os Sistemas de Integrados de Produção Agropecuária (SIPA) são apontados como uma alternativa agrícola sustentável para maximizar o uso do solo, com grande possibilidade de adoção pelos agricultores (CARVALHO et al., 2005). O SIPA utiliza-se das interações positivas entre planta e animal, o que resulta em melhorias ambientais e viabilidade econômica das propriedades rurais (CARVALHO et al., 2010).

Este sistema é planejado para explorar produtos e propriedades emergentes oriundos das interações entre solo, planta, animal e atmosfera (ANGHINONI, 2013), e baseia-se na diversificação, rotação, consorciação e/ou sucessão das atividades da agricultura e da pecuária dentro da propriedade rural (KLUTHCOUSKI et al., 1991). Este manejo, se aplicado de forma harmônica, traz benefícios para ambas as atividades, onde o solo é explorado durante o ano inteiro, o que favorece o aumento na oferta de grãos, de carne e de leite a um baixo custo (ALVARENGA e NOCE, 2005). Além dos benefícios econômicos, o SIPA é associado à melhoria dos atributos físicos e químicos do solo e redução de pragas e doenças (ASSMANN et al., 2015; CECAGNO et al., 2016; SILVA et al., 2014b; WESP et al., 2016).

O SIPA traz inúmeros benefícios ao solo, quando bem manejado. O solo atua como mediador dos vários processos aos quais o sistema é submetido, incorporando nutrientes e energia oriundos de uma diversidade vegetal, associados a novas vias de fluxo de nutrientes e água de origem animal (ANGHINONI, 2013). Como consequência tem-se a melhoria de seus aspectos físicos, químicos e biológicos, além do aumento de produtividade das culturas de interesse agrícola e das pastagens, e renda adicional

com a venda dos animais resultando em melhoria da qualidade de vida do produtor (BERRY et al., 2003; VILELA et al., 2008). Aliado a isso está a facilidade do sistema se adaptar a qualquer tamanho de propriedade, desde que as características de solo não apresentem restrições.

O SIPA pode afetar direta e intensamente a biologia do solo através de vários processos. Destes, cabe citar a alteração na quantidade e qualidade dos resíduos aportados (SILVA et al., 2014a), o aporte diversificado de resíduos vegetais (SOUZA et al., 2010), as modificações na direção, magnitude e composição dos fluxos de nutrientes (ANGHINONI, 2013) e as alterações nos atributos físicos do solo (CECAGNO et al., 2016). Poucos foram os estudos realizados abordando as comunidades biológicas neste tipo de sistema. Para comunidades microbianas foi observada resposta rápida às mudanças no pastejo (LE ROUX et al., 2007), onde a intensidade de pastejo moderada pode apresentar aumentos tanto na atividade quanto na diversidade microbiológica, se comparadas a áreas de intenso ou nenhum pastejo (ZHOU et al., 2010). Para a fauna de invertebrados edáficos, o SIPA parece favorecer a manutenção da diversidade da fauna invertebrada (PORTILHO et al., 2011) e beneficiar um ambiente edáfico biologicamente mais ativo, se comparado aos sistemas convencionais (SILVA et al., 2011). Resultados na região subtropical do Brasil acerca da fauna e da microbiota do solo relacionados ao SIPA ainda são escassos.

1.2 COMUNIDADE DE ORGANISMOS EDÁFICOS COMO BIOINDICADORES DE QUALIDADE DO SOLO

A qualidade do solo afeta diretamente a produtividade e a sustentabilidade dos sistemas agrícolas. Pode-se representar a qualidade por um conjunto de parâmetros obtidos no solo, através de suas propriedades físicas, químicas e biológicas (LARSON e PIERCE, 1994), e a escolha dos parâmetros a serem utilizados depende do objetivo da avaliação a ser feita (GARDI et al., 2002). Os indicadores biológicos de qualidade do solo têm sido utilizados com frequência em diversos estudos para avaliar a qualidade em diferentes sistemas de plantio (BARTZ et al., 2013; CLUZEAU et al., 2012; FLOCH et al., 2011; VASCONCELLOS et al., 2013). As principais características exigidas para ser um bom indicador de qualidade do solo são apresentar estreita relação com funções do solo, sensibilidade e rápida resposta às

mudanças no meio (DORAN e ZEISS, 2000). Pode-se considerar que a principal vantagem dos bioindicadores seja o fato de eles constituírem o atributo vivo do solo, apresentando, portanto, mais rápida resposta a variação no ecossistema se comparados com os atributos químicos ou físicos.

O ambiente edáfico representa um hábitat natural que comporta uma grande e diversa comunidade de seres vivos, que vão desde microrganismos até um grande grupo de espécies de macroinvertebrados. Estes organismos influenciam o meio no qual eles vivem, pois são responsáveis por funções ambientais importantes, atuando diretamente sobre a decomposição da matéria orgânica, a ciclagem de nutrientes, e a estruturação do solo (LAVELLE et al., 1993; LAVELLE e SPAIN, 2002). Da mesma forma, os organismos são influenciados pelo meio. A intensidade do uso do solo e o tipo e a qualidade da cobertura vegetal alteram de forma direta e significativa a biodiversidade edáfica (ROSA et al., 2015). A simplificação de hábitat gera mudanças no aporte de resíduos, na oferta de recursos, alterações na temperatura e umidade que podem levar a mudanças drásticas na composição das comunidades biológicas, levando ao seu declínio (BEDANO et al., 2016; LITTLE et al., 2013; PORTILHO et al., 2011).

Os microrganismos do solo, constituídos por bactérias, fungos, algas e protozoários, apresentam ativa participação nos processos de mineralização e imobilização temporária de nutrientes, ciclagem de nutrientes, fluxos de energia e transformação da matéria orgânica no solo (FERREIRA et al., 2011; KOTROCZÓ et al., 2014; SPOHN et al., 2016). Sua atividade é facilmente influenciada por fatores que alteram o solo, como manejo adotado, cobertura vegetal e tipo de fertilização (CARRERA et al., 2007; SHARKHUU et al., 2016), estágio do desenvolvimento de plantas e uso de pesticidas (FERREIRA et al., 2009; FERREIRA et al., 2008). Portanto, os microrganismos são considerados indicadores sensíveis da qualidade do solo (FERREIRA et al., 2011). Assim como as comunidades microbianas, os indicadores microbianos também se apresentam como boas alternativas de avaliação da qualidade do solo (LISBOA et al., 2012). Características como biomassa microbiana do solo (BMS), enzimas extracelulares e taxa de respiração basal do solo podem ser utilizadas como bioindicadoras de qualidade do solo, já que elas são ligadas intimamente com propriedades edáficas importantes, dentre as quais cabe citar os teores de matéria orgânica, a ciclagem de nutrientes e algumas propriedades físicas e químicas do solo (MOSCATELLI et al., 2012; VASCONCELLOS et al., 2013).

A comunidade de invertebrados do solo é diretamente relacionada à transformação, decomposição e liberação de nutrientes a partir de resíduos orgânicos, à estruturação do solo, a criação de bioporos e ao revolvimento do solo, incorporando matéria orgânica ao longo do perfil (BARTZ et al., 2014; ROVEDDER et al., 2009; STEFFEN et al., 2007). Apesar desta relevância ecológica, a atividade e diversidade dos organismos edáficos é sensível ao tipo de uso do solo adotado e pode ser prejudicada em sistemas mal manejados (BARETTA et al., 2014; BARTZ et al., 2014; ROSA et al., 2015). Assim, a abundância, a diversidade e a atividade da fauna edáfica podem fornecer indicativos úteis sobre a qualidade do solo (BARETTA et al., 2014; LAVELLE et al., 2006).

Várias são as opções de bioindicadores de qualidade do solo, porém estudos avaliando a comunidade biológica e a correlacionando com as variáveis obtidas em solos sob sistemas de SIPA são raramente efetuados. Sabe-se do potencial benéfico que o SIPA apresenta sobre os atributos químicos e físicos do solo, e da estreita ligação das comunidades edáficas com estes. Logo, torna-se necessária uma avaliação mais aprofundada do comportamento destas comunidades em solos submetidos ao SIPA.

2 HIPÓTESES E OBJETIVOS

2.1 HIPÓTESES

Em um sistema integrado de produção agropecuária quanto maior for a intensidade do pastejo, maior é a redução da atividade e da diversidade dos organismos do solo;

As áreas com maior intensidade de pastejo apresentam menor qualidade física do solo e menor cobertura vegetal, o que resulta em menor atividade e diversidade da biota edáfica;

Os atributos biológicos do solo são adequados indicadores de qualidade ambiental dos sistemas integrados de produção agropecuária.

2.2 OBJETIVOS

2.2.1 Objetivo Geral

Avaliar as comunidades de micro, meso e macrorganismos do solo após 15 anos de um SIPA manejado sob diferentes intensidades de pastejo e verificar se estes parâmetros biológicos podem ser utilizados como bioindicadores de qualidade neste sistema.

2.2.2 Objetivos Específicos

Determinar a biomassa e a atividade dos microrganismos do solo após 15 anos de pastejo com diferentes intensidades em um sistema integrado de produção agropecuária;

Avaliar a abundância e diversidade dos meso e macrorganismos epiedáficos em um SIPA de longa duração manejado sob diferentes intensidades de pastejo;

Realizar o levantamento das principais características físicas e químicas do solo e de cobertura vegetal após 15 anos de pastejo com diferentes intensidades em um sistema integrado de produção agropecuária;

Estabelecer relações entre os atributos físicos, químicos e de cobertura do solo com os atributos biológicos do solo.

26 the treatment with the lowest grazing height, resulting in decreased soil moisture and
27 microbial biomass and activity.

28 Keywords: microorganism, metabolic quotient, bioindicator, integrated crop-livestock
29 systems, sustainability.

30

31 **1. Introduction**

32 The introduction of animal grazing to crop production areas may improve soil
33 physical, chemical and biological properties (Berry *et al.*, 2003; Vilela *et al.*, 2008).
34 The presence of animals results in the incorporation of feces and urine into the soil,
35 resulting in higher forage plant biomass, root activity and dry mass per growth
36 season under grazing conditions. All this contributes to create a favorable
37 environment for the growth and activity of soil microorganisms.

38 However, inadequate management of integrated crop-livestock systems (ICLs)
39 due to an increased grazing load on pastures is often observed (Neves Neto *et al.*,
40 2013). Intensive grazing may outweigh the benefits of ICLs and have detrimental
41 effects on soil. High grazing load has been reported to change the soil temperature
42 and moisture regime (Klein *et al.*, 2005), and intensive grazing was observed to
43 result in decreased soil vegetation cover and organic carbon concentration as well as
44 increased topsoil compaction (Kölbl *et al.*, 2011). All these changes directly or
45 indirectly affect microbial biomass and activity, and these parameters can therefore
46 serve as bioindicators of an ICL's ecosystem quality and balance.

47 Recent studies indicate that soil microbial communities may be affected by
48 grazing. The introduction of grazing to croplands was observed to result in increased
49 microbial activity and soil basal respiration rates in Mongolia (Sharkhuu *et al.*, 2016)
50 and the United States (Adewopo *et al.*, 2015). In China, grasslands were observed to

51 present a more stable C-CO₂ emission rate throughout the day than regenerated
52 forests or farmlands (Liu *et al.*, 2016).

53 Although the ICL is increasingly acknowledged as an economic, social and
54 environmental alternative to increase food production in several countries, few
55 studies have focused on the long-term effects of intensive grazing on the dynamics of
56 soil microbial communities, especially under subtropical soil and climate conditions.
57 An experiment was performed to test the hypothesis that intensive grazing in a long-
58 term ICL decreases soil microbial biomass and activity. The aim of the present study
59 was to investigate the effects of 15 years of intensive grazing in an integrated
60 soybean-beef cattle system on soil microbial biomass and respiration, identifying
61 which edaphic factors most affect these parameters.

62

63 **2. Materials and methods**

64 2.1. Experimental site and conduction of the experiment

65 The study was performed as part of ongoing research conducted since 2001
66 by the Integrated Crop-Livestock Systems Research Group (Grupo de Pesquisa em
67 Sistema Integrado de Produção Agropecuária) of the Federal University of Rio
68 Grande do Sul (Universidade Federal do Rio Grande do Sul – UFRGS), in an area
69 comprising approximately 23 ha, located in the municipality of São Miguel das
70 Missões, Rio Grande do Sul, Brazil (29°03'10"S, 53°50'44"W). The soil is clayey (540
71 g kg⁻¹ clay for a soil layer 0–20 cm deep) and classified as an Oxisol (Rhodic
72 Hapludox - Soil Survey Staff, 1999). The climate is subtropical with hot and humid
73 summers (type Cfa), according to the Köppen climate classification.

74 Before the experiment began, the area had been cultivated under no tillage
75 since 1993, with black oat (*Avena strigosa* Schreb) in winter (only for soil cover) and

76 soybean [*Glycine max* (L.) Merr.] in summer. From the beginning of the experiment
77 (2001), until the present, the area has been continuously cultivated with soybean in
78 summer (November to April) for grain production, under no tillage, and the cultivation
79 of black oat + Italian ryegrass (*Lolium multiflorum* Lam.) was initiated in winter (May
80 to October) for pasture, with four different grazing heights (treatments).

81 The treatments consisted of four grazing intensities: intense (10-cm grazing
82 height), moderate (20- and 30-cm grazing heights), light (40-cm grazing height) and
83 no grazing (control plots). A randomized block experimental design was used, with
84 three replicates per treatment. The size of the plots with grazing varied from 0.9 to
85 3.6 ha. All plots had continuous grazing, a variable stocking rate and three tester
86 animals and grazer animals as needed (200 kg average initial live weight), resulting
87 in the different grazing heights tested.

88 Fertilization was based on the soil analysis and performed according to
89 technical recommendations. In pastures, nitrogen fertilization was applied to obtain a
90 yield between 4.0 and 7.0 t ha⁻¹ of pasture dry matter. In soybean plantations,
91 phosphorus and potassium fertilization was applied to obtain a yield of 4.0 t ha⁻¹
92 (Assmann *et al.*, 2015; Schuster *et al.*, 2016).

93 2.2. Samplings and analyses

94 For the microbiological analyses, two samplings were performed immediately
95 after the end of the grazing season (beginning of November 2014 and 2015) and two
96 immediately after the soybean harvest (beginning of May 2015 and 2016). Five soil
97 samples were collected from each plot, from the corners and center of a 36 x 60-m
98 virtual quadrat placed in the center of the plot (coordinates obtained using a precision
99 real time kinematic global positioning system). For all samplings, collections were
100 always performed at the same points, from the 0–10 cm soil layer, using a soil-

101 sampling auger. Soil samples were placed in plastic bags, stored in styrofoam boxes
102 with ice and transported to the laboratory, where they were sieved and stored at 4°C
103 for up to one week until analyzed.

104 Microbial biomass carbon (MBC) was determined using the fumigation-
105 extraction method (De-Polli and Guerra, 1997), and carbon (C) was quantified using
106 a total organic carbon analyzer (TOC-L Shimadzu). The soil basal respiration rate
107 was determined by respirometry (Anderson and Domsch, 1978), consisting of soil
108 incubation for 21 days in the dark at 26°C. The metabolic quotient for CO₂ (qCO₂)
109 was calculated as the ratio between MBC and the accumulated respiration rate
110 (Anderson and Domsch, 1993).

111 Samples for physical and chemical soil analyses were collected from the 0–10
112 cm soil layer in November 2014 and April 2015 at the same five points in each plot,
113 as described above. The following soil chemical parameters were determined: total
114 organic carbon (Walkley–Black); pH (water 1:1); P and K (Mehlich-1); Ca, Mg and Al
115 (KCl 1 mol L⁻¹); V (% base saturation); and H+Al (Toledo *et al.*, 2012). The following
116 soil physical parameters were determined: total porosity, macroporosity and
117 microporosity, using a tension table (Embrapa, 1997); soil density, using the
118 volumetric ring method (Blake and Hartge, 1986); and gravimetric moisture (Table 1).

119 Litter and plant shoot samples were collected from three points in each plot,
120 on the same days as the soil collections, for microbiological analyses. Plant material
121 was dried in a forced air ventilation oven at 65°C until constant weight was achieved.
122 Rainfall and air temperature data were collected for two months for each sampling
123 season, using a meteorological station situated in the experimental area (Figure 1).

124 2.3. Statistical analyses

125 An analysis of variance (ANOVA) was performed, followed by a Duncan test
126 when significant differences were found, at $p \leq 0.05$, using the SASM-agri (version
127 3.2.4) software. Regression analyses were performed to analyze the relationship
128 between MBC and the soil basal respiration rate, for the cycle following the grazing
129 season and following soybean cultivation. The relationships between microbiological
130 parameters and soil moisture, litter and plant biomass, for each sampling, were
131 analyzed using the Pearson's correlation coefficient ($p \leq 0.05$), using the SigmaPlot
132 (version 11.0) software.

133 **3. Results**

134 C-CO₂ production was higher for the treatments without grazing (control) and
135 with higher grazing heights, for all samplings (Figure 2), indicating that intensive
136 grazing decreased soil microbial activity. On average, C-CO₂ production was 16, 19,
137 26 and 30% lower for the treatment with the 10-cm grazing height than with the 20-,
138 30- or 40-cm grazing height or without grazing, respectively (Figure 2). The treatment
139 with the 40-cm grazing height presented higher C-CO₂ production than the other
140 grazing treatments, exhibiting values similar to those observed for the control
141 treatment, for most seasons.

142 Similar to basal respiration, the MBC content also indicated detrimental effects
143 of intense grazing on soil microorganism biomass. Except for the sampling performed
144 in May 2015, MBC was highest for the treatments without grazing and with the
145 highest grazing heights (Figure 2). On average, MBC was 2, 4, 9 and 18% lower for
146 the treatment with the 10-cm grazing height than with the 20-, 30- or 40-cm grazing
147 height or without grazing, respectively (Figure 2). A linear and positive correlation
148 between MBC and basal respiration was observed for the samplings following
149 grazing ($R^2=0.88$; Figure 3), but not for the samplings following the soybean harvest,

150 for which a low coefficient of determination was observed ($R^2=0.43$). This was
151 possibly due to the May 2015 sampling, which presented low basal respiration rates
152 and high MBC.

153 The qCO_2 was low for all treatments and seasons, with a maximum value of
154 $0.29 \text{ mg C-CO}_2 \text{ kg}^{-1} \text{ mg MBC kg}^{-1}$ (Figure 2). The differences between treatments
155 were therefore also very small, which contributed to the absence of an observable
156 trend between treatments. The low qCO_2 values indicate that the microbial
157 community in the studied ICL was stable, even under intense grazing.

158 Soil moisture was highest for the control treatment (without grazing) for all
159 samplings, although not statistically significantly different for the first samplings, and
160 tended to increase with increasing grazing height (Figure 4). This seemed to be due
161 to a higher shoot and litter dry mass and better soil physical conditions, observed for
162 the treatments without grazing and with the highest grazing heights. Soil density was
163 lower and soil porosity was higher for these treatments (Table 1). The highest density
164 and lowest porosity were observed for the 10-cm grazing height, which resulted in
165 lower soil moisture (Figure 4).

166 For most samplings, the microbial basal respiration or MBC content were only
167 significantly correlated with soil moisture and litter and shoot dry mass (Table 2).
168 Basal respiration was positively and significantly correlated with shoot and litter dry
169 mass for all samplings and with soil moisture for the May and November 2015
170 samplings. MBC was positively correlated with litter dry mass in November 2014 and
171 with moisture, litter and shoot dry mass in November 2015.

172 **4. Discussion**

173 The present data were obtained from four samplings, performed over two
174 years, as part of an experiment conducted over a 15-year period, with the same

175 treatments, in a 23-ha area, supported by a very large soil, plant, animal and
176 meteorological data set collected continuously during the duration of the experiment.
177 This experiment therefore offers an excellent opportunity to study soil microbiology in
178 an ICL and decreases the global knowledge gap on this subject.

179 Many farmers in southern Brazil currently use non-integrated crop production
180 systems. In winter, 86% of the agricultural area is cultivated with cover plants for
181 straw production, to be used in the no-tillage sowing of soybean in summer (Conab,
182 2011). This concentrates farmers' entire income on soybean, which is a great risk,
183 considering weather and market instabilities. In turn, oat and Italian ryegrass winter
184 pastures are highly productive; they can be used for animal grazing and still produce
185 enough plant biomass for no-tillage sowing in the summer (Kichel and Miranda,
186 2006; Vilela *et al.*, 2008).

187 The introduction of cattle grazing in winter is an attempt to increase and
188 stabilize farmers' incomes through the intensification of soil use (Berry *et al.*, 2003).
189 However, whether introducing grazing can be detrimental to the soil is frequently
190 questioned. Grazing may increase soil microbial activity and biomass, as it results in
191 higher residue input and diversity, due to the incorporation of cattle urine and manure
192 (Clegg, 2006; Lin *et al.*, 2009). In addition, forage plant root activity and growth are
193 promoted by the defoliation caused by grazing, which results in an increased
194 rhizodeposition turnover rate and fine root decomposition (Papatheodorou *et al.*,
195 2008; Hewins *et al.*, 2016). In turn, grazing may be detrimental to microbial activity
196 because it increases topsoil compaction and decreases soil porosity, decreasing soil
197 aeration and water infiltration, and removes part of the plant biomass, altering soil
198 temperature and moisture (Jia *et al.*, 2006; Souto *et al.*, 2008).

199 The aim of the present study was to understand the effects of different grazing
200 heights on soil microbial biomass and activity, with a focus on ICL sustainability. Soil
201 microbial biomass and activity were chosen because they are highly sensitive
202 bioindicators of environmental quality, and because high soil biological activity is a
203 requirement for the sustainability of any production system. In biologically active
204 soils, organic residue degradation, nutrient mineralization, soil organic matter
205 formation, soil aggregation, biological control, plant hormone production and other
206 processes occur in adequate levels, contributing to increased crop productivity and
207 environmental conservation.

208 The present results indicate that intense grazing decreases soil microbial
209 activity and biomass in an ICL. Overall, soil respiration and MBC were higher in
210 treatments without grazing and with moderate to light grazing and were directly
211 correlated to soil moisture and litter and shoot biomass. Soil moisture is one of the
212 primary factors affecting soil microbial community composition and therefore soil
213 respiration (Jia *et al.*, 2006; Chen *et al.*, 2015; Liu *et al.*, 2016). Previous studies have
214 indicated that the seasonal variation of C-CO₂ production is predominantly governed
215 by soil moisture and temperature (Risch and Frank, 2010).

216 Soil moisture was lower for the intense grazing treatment in all samplings,
217 coinciding with lower plant cover, lower soil porosity and higher soil density, resulting
218 in an environment with lower microbial activity and biomass. In addition, higher
219 removal of shoot biomass results in decreased soil protection against high levels of
220 solar radiation and consequently higher evaporation (Gong *et al.*, 2014).

221 The present results showed that the higher the grazing pressure is, the lower
222 the soil plant and litter cover and the lower the basal respiration rate and MBC. A
223 higher removal of shoot biomass, and consequent decrease in litter, results in lower

224 soil protection against high levels of solar radiation and higher evaporation (Gong *et*
225 *al.*, 2014), which causes the habitat conditions to be inadequate for microbial activity.

226 Overall, C-CO₂ fluxes were higher in areas without grazing or with moderate to
227 light grazing. Peri *et al.* (2015) observed up to 30% higher microbial activity in areas
228 under moderate grazing than under intense grazing. The exclusion of grazing also
229 increases microbial activity (Prem *et al.*, 2014). However, C cycling was observed to
230 slow following 10 years of grazing exclusion (Medina-Roldán *et al.*, 2012).

231 For all treatments, the lowest MBC values were observed for the first and last
232 sampling periods, coinciding with the highest and lowest rainfall quantities,
233 respectively. Soil O₂ supply may decrease in periods of high rainfall, which is directly
234 reflected in the amount and activity of aerobic microorganisms (Souto *et al.*, 2008). In
235 turn, low rainfall generates water stress and decreases soil diffusion, which is
236 detrimental to the microbial community (Manzoni *et al.*, 2012; Sharkhuu *et al.*, 2016).

237 MBC typically corresponds to 2 to 4% of the total soil organic carbon stock
238 (TOC) and is the organic matter fraction that is most sensitive to changes resulting
239 from management (Gama-Rodrigues, 1999). In the present study, the contribution of
240 MBC to TOC varied between samplings, averaging 2% for the first and fourth
241 samplings, and 4% for the second and third samplings, indicating satisfactory
242 microbial growth in the study area. The high MBC observed in the present study is in
243 accordance with the MBC values between 0.9 and 1.8 g kg⁻¹ observed in previous
244 long-term grazing experiments (Prem *et al.*, 2014; Spohn *et al.*, 2016; Stevenson *et*
245 *al.*, 2016). Except for the May 2016 sampling, no significant differences in MBC were
246 observed between the moderate and light grazing treatment, indicating that the
247 microbial communities were well adapted to the ICL. Pasture ecosystems with a high
248 organic matter input rate promote the growth and activity of microbial communities,

249 through the continuous inclusion of plant and animal residues, and promote root
250 turnover, therefore increasing soil respiration rates (Peri *et al.*, 2015).

251 The qCO_2 was low for all treatments (lower than $0.3 \text{ mg C-CO}_2 \text{ kg}^{-1}$), indicating
252 a low level of physiological stress in the microbial community. A high qCO_2 indicates
253 that adverse or stressful conditions are present in the microbial population and
254 therefore that the soil organic carbon is being inadequately managed (Anderson and
255 Domsch, 1993). Studies evaluating qCO_2 under several types of management
256 reported $1.9 \text{ mg C-CO}_2 \text{ kg}^{-1}$ for pastures (Stevenson *et al.*, 2016) and 17.8 mg C-CO_2
257 kg^{-1} for soybean plantations (Zilli *et al.*, 2008). ICL characteristics, such as the
258 frequent input of organic animal and plant residues, permanent soil cultivation and a
259 long experimental duration, contribute to this stability.

260 Despite the overall low stress level of the microbial community observed for all
261 treatments, decreased soil microbial biomass and activity were observed for lower
262 grazing heights. The detriment to soil microorganisms caused by cattle grazing was
263 minimized at the highest grazing heights. The taller living and dead plant biomass in
264 these treatments led to better soil physical conditions, resulting in higher soil
265 moisture and therefore higher microbial activity and biomass. These results
266 emphasize the need for adequately adjusting the animal load on pastures, to not
267 compromise ICL sustainability.

268 **5. Conclusions**

269 Soil microbial respiration and microbial biomass carbon content were higher
270 without grazing and at the highest grazing heights.

271 Negative effects on the physical characteristics of soil were observed for the
272 lowest grazing height, resulting in lower soil moisture content, microbial activity and
273 biomass.

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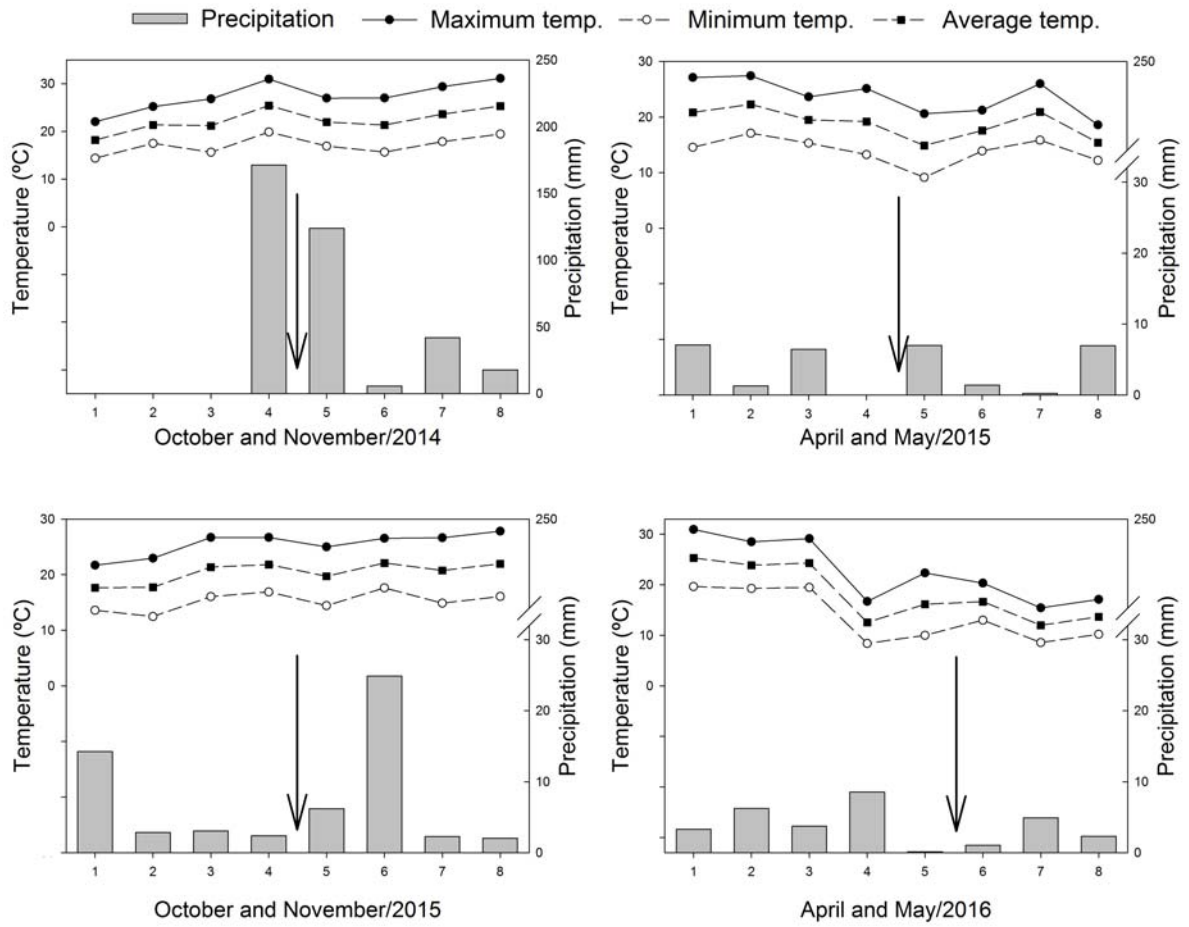
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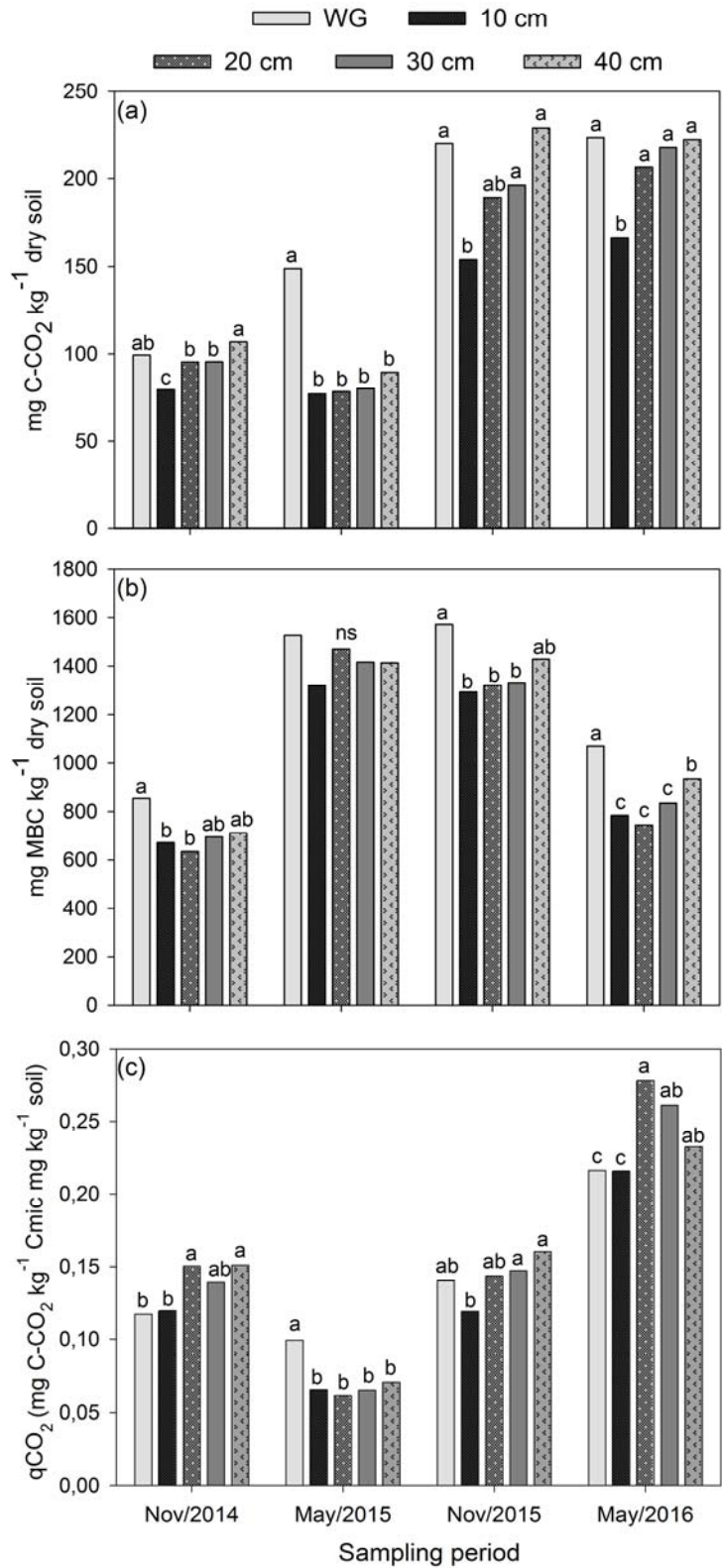
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381 Figure 1: Rainfall and minimum, maximum and mean temperatures for all sampling
 382 periods. The arrow indicates the soil sampling date (11/05/2014; 04/29/2015;
 383 11/01/2015; 05/05/2016).

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387 Figure 2: Basal respiration (a), microbial biomass carbon content (b) and metabolic

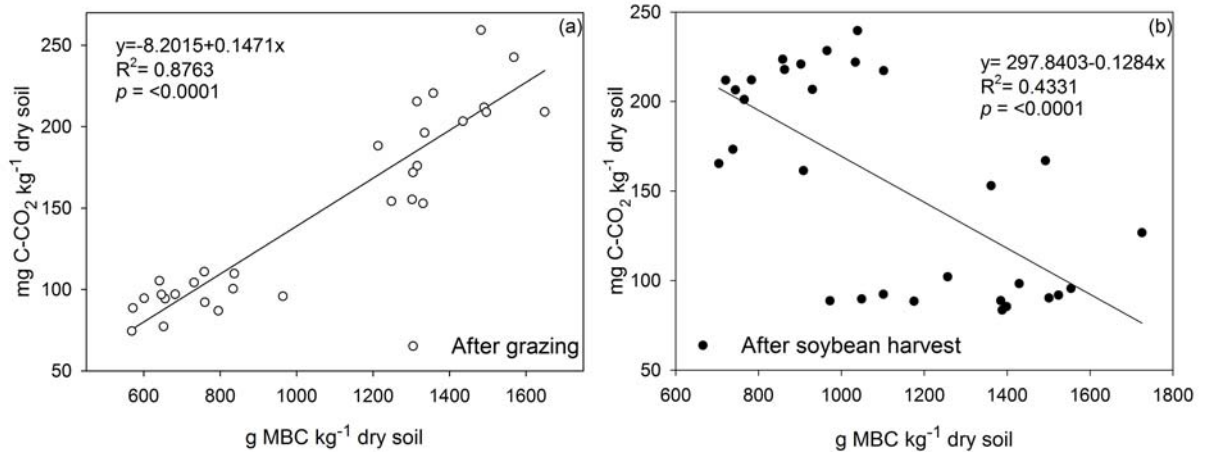
388 quotient (c) at the 0–10 cm soil layer, following the cattle grazing season (Nov/2014

389 and Nov/2015) or soybean harvest (May/2015 and May/2016), for different
390 treatments with different grazing heights (10, 20, 30 and 40 cm) or without grazing
391 (WG), in a 15-year-old integrated soybean-beef cattle production system. Values are
392 means of 15 replicates per treatment. Means followed by different letters were
393 significantly different according to the Duncan test at $p \leq 0.05$.

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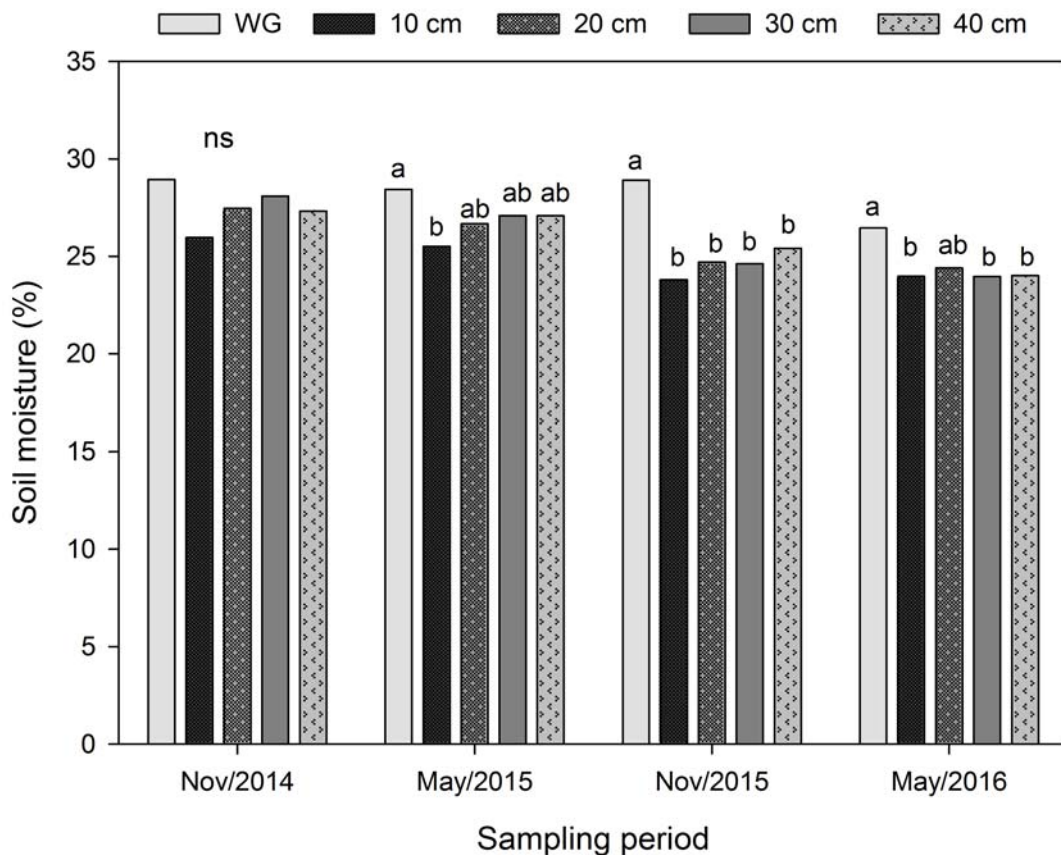
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398 Figure 3: Linear regression between microbial biomass carbon and soil basal
399 respiration rate for samplings performed following the cattle grazing season (a) or
400 soybean harvest (b) in a 15-year-old integrated soybean-beef cattle production
401 system with different grazing intensities (n=30).



402

403 Figure 4: Gravimetric soil moisture at the 0–10 cm soil layer following the cattle
 404 grazing season (Nov/2014 and Nov/2015) or soybean harvest (May/2015 and
 405 May/2016), for different treatments with different grazing heights or without grazing
 406 (WG), in a 15-year-old integrated soybean-beef cattle production system. Values are
 407 means of 15 replicates per treatment. Means followed by different letters were
 408 significantly different according to the Duncan test at $p \leq 0.05$.

409

410 Table 1: Soil physical and chemical characteristics at the 0–10 cm soil layer and
 411 plant cover following the cattle grazing season (November) or soybean harvest
 412 (May), for different treatments with different grazing heights or without grazing (WG),
 413 in a 15-year-old integrated soybean-beef cattle production system. Values are means
 414 of 5 replicates per treatment.

Treat.	SD ¹	Ma ²	Mi ³	pH	P	K	TOC ⁴	V ⁵	Al	Ca	Mg	H+Al	LDM ⁶		SDM ⁷	
	g cm ⁻³	cm ³ cm ⁻³	cm ³ cm ⁻³		mg dm ⁻³	g kg ⁻¹	g kg ⁻¹	%		cmol _c dm ⁻³	cmol _c dm ⁻³	cmol _c dm ⁻³	2014	2015	2014	2015
	After grazing season												2014	2015	2014	2015
WG	1.26	0.13	0.42	4.3	13.1	219.3	- ⁸	27.1	1.4	3.0	1.2	13.6	6.6	5.1	4.6	7.3
10 cm	1.40	0.09	0.40	4.8	9.0	187.4	-	48.5	0.8	4.1	1.7	6.9	1.3	1.0	1.8	0.9
20 cm	1.37	0.09	0.41	4.9	10.7	200.4	-	50.6	0.6	4.0	1.8	6.4	3.0	1.8	3.3	2.6
30 cm	1.33	0.12	0.42	5.0	9.2	163.4	-	52.5	0.6	4.5	1.9	6.4	3.1	2.1	3.4	3.3
40 cm	1.32	0.14	0.38	4.8	13.6	200.6	-	48.6	0.5	4.4	1.8	7.2	3.5	4.0	4.8	4.1
	After soybean harvest												2015	2016	2015	2016
WG	1.34	0.09	0.45	4.3	19.5	222.6	23.1	36.4	1.3	3.9	1.5	11.2	9.2	8.5	0 ⁹	0
10 cm	1.37	0.09	0.41	4.6	9.7	136.3	22.2	44.6	1.1	4.3	2.0	8.5	4.1	4.3	0	0
20 cm	1.36	0.11	0.41	4.7	13.6	181.2	23.3	47.5	1.0	4.7	2.2	8.2	6.3	6.0	0	0
30 cm	1.36	0.09	0.43	4.8	9.4	148.3	22.8	49.0	0.7	4.6	2.1	7.5	6.6	6.0	0	0
40 cm	1.33	0.10	0.43	4.6	12.6	173.8	22.3	45.7	0.8	4.7	1.9	8.4	6.9	8.0	0	0

415 ¹Soil density; ²Macroporosity; ³Microporosity; ⁴Total organic carbon; ⁵Base saturation; ⁶Litter
 416 dry mass; ⁷Shoot dry mass; ⁸Not sampled; ⁹After the soybean harvest, no plants were
 417 growing in the area.

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432 Table 2: Pearson correlation between soil basal respiration (C-CO₂) or microbial
 433 biomass carbon (MBC) and soil moisture, litter dry mass (LDM) and shoot dry mass
 434 (SDM), following the cattle grazing season (Nov/2014 and Nov/2015) or soybean
 435 harvest (May/2015 and May/2016), in a 15-year-old integrated soybean-beef cattle
 436 production system with different grazing heights (n=15).

Sampling period	C-CO ₂			MBC		
	Soil moisture	LDM	SDM	Soil moisture	LDM	SDM
Nov/14	0.23	0.50*	0.77**	0.26	0.62**	0.35
May/15	0.63**	0.51*	- ¹	0.41	0.23	-
Nov/15	0.58*	0.70**	0.79***	0.75***	0.70**	0.77***
May/16	0.21	0.51*	-	0.47	0.45	-

437 ¹After the soybean harvest, no plants were growing in the area. * $p < 0.05$; ** $p < 0.01$; *** $p < 0.001$.

1 **TITLE**

2 **Effect of long-term grazing intensification on epiedaphic faunal diversity**
3 **in an integrated crop-livestock****

4 **AUTHORS**

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6 Antonioli; Dilmar Baretta; Ibanor Anghinoni; Rodrigo Josemar Seminoti
7 Jacques

8 **ABSTRACT**

9 Epiedaphic fauna is responsible for various key processes in the maintenance
10 of soil and environmental quality. Intensification of grazing in integrated crop-
11 livestock (ICL) systems may lead to a loss of diversity in soil organisms and
12 compromise the sustainability of the system. This study aimed to identify the
13 effects of 15 years of grazing intensification in an integrated soybean-beef cattle
14 system on the diversity of epiedaphic fauna and to assess which edaphic
15 conditions most affect the community of these organisms. The assessments
16 were performed in an experiment conducted beginning in 2001 in a 23-ha area
17 cropped with *Glycine max* for grain production in the summer and *Avena*
18 *strigosa* + *Lolium multiflorum* for continuous cattle grazing in the winter. The
19 treatments consisted of grass heights of 10, 20, 30, and 40 cm and controls
20 without grazing. Sampling was performed in four collection periods from 2014 to
21 2016: two after grazing and two after soybean harvest. Epiedaphic faunal
22 diversity was assessed using pitfall traps and chemical, physical, and soil cover
23 variables for correlations. Grazing intensification for 15 years decreased the
24 epiedaphic faunal abundance, richness, and diversity. The Collembola, Acari,
25 Coleoptera, Araneae, Orthoptera, Dermaptera, and Hymenoptera groups were

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26 the most sensitive to the different grazing intensities. Vegetation cover and soil
27 moisture were the factors that most affect the epiedaphic organisms in an
28 integrated soybean-beef cattle system.

29 KEYWORDS: edaphic invertebrates; soil macrofauna; soil mesofauna;
30 integrated crop-livestock system; soil physical quality.

31

32 **1. INTRODUCTION**

33 - Integrated crop-livestock (ICL) systems are based on the diversification,
34 rotation, and combination and/or succession of crop and livestock production
35 activities in farms (Kluthcouski et al. 1991). This management system, if applied
36 correctly, brings benefits to both activities. The soil is exploited for nearly the
37 entire year, which favors an increase in supplies of grain, meat, and/or milk and
38 reduces production costs (Alvarenga and Noce 2005).

39 - ICL may improve the soil quality and resilience capacity, in addition to
40 providing economic benefits (Salton et al. 2014). Studies have shown that this
41 system can increase the carbon, nitrogen (Assmann et al. 2015), potassium,
42 and phosphorus (Silva et al. 2014) cycling efficiency without damaging the soil
43 physical quality when managed at moderate grazing intensity with grass heights
44 of 20 to 30 cm (Cecagno et al. 2016). However, fewer research studies have
45 focused on ICL biological attributes, particularly with long-term experimental
46 protocols.

47 - Epiedaphic fauna consists of the invertebrates inhabiting the soil surface. The
48 activities of these organisms improve agricultural productivity and environmental
49 quality, resulting in the incorporation of soil surface residues, increased organic
50 matter and plant nutrient availability, improved aggregation, biopore opening,

51 increased water infiltration and aeration, biological control, and other factors
52 (Ferreira et al. 2007; Lavelle 1997; Rovedder et al. 2009). However, the
53 management adopted in ICL may directly and intensely affect the epiedaphic
54 fauna because these organisms are sensitive to soil chemical and physical
55 changes and to changes in the vegetation and in the quantity and diversity of
56 the residues deposited on the soil (Greenwood and McKenzie 2001; Souza et
57 al. 2010).

58 - A problem commonly observed in ICL is the excessive stocking rate of
59 grasslands, which results in reduced living plant and litter biomass availability
60 (Kölbl et al. 2011), topsoil compaction and reduced water infiltration (Cecagno
61 et al. 2016), and altered soil moisture and temperature regimes (Klein et al.
62 2005; Neves Neto et al. 2013), among other damage. Therefore, inadequate
63 ICL management may offset its benefits and reduce soil biodiversity.

64 - Considering the importance of epiedaphic organisms to the sustainability of
65 agro-ecosystems and the sensitivity of these organisms to management
66 systems, this study tested the hypothesis that long-term grazing intensification
67 in an ICL reduces the diversity of the main groups of soil invertebrates.
68 Accordingly, we aimed to assess the effects of 15 years of grazing
69 intensification in an integrated soybean-beef cattle production system on the
70 diversity of the epiedaphic fauna and determine which edaphic conditions most
71 affect the community of these organisms.

72 **2. MATERIALS AND METHODS**

73 2.1. Experimental site and procedures

74 - The study was performed as part of an experiment conducted beginning in
75 2001 in an area of approximately 23 ha by the Integrated Agricultural

76 Production System Research Group of the Federal University of Rio Grande do
77 Sul (Universidade Federal do Rio Grande do Sul - UFRGS). The area is located
78 in the municipality of São Miguel das Missões/ Rio Grande do Sul State/ Brazil
79 (29°03'10" South, 53°50'44" West) and has a clayey soil (540 g kg⁻¹ clay in the
80 0- to 20-cm layer) classified as Oxisol (Rhodic Hapludox - Soil Survey Staff
81 1999). The climate is subtropical, with a hot and humid summer (Cfa),
82 according to the Köppen climate classification. Before 2001, the area was
83 planted with black oat (*Avena strigosa* Schreb) in winters (only for soil cover)
84 and soybean (*Glycine max* (L.) Merr.) in summers in a no-till crop production
85 system. After the experiment had been established in the area, soybean
86 planting was maintained for grain production in the summers (November to
87 April), although the black oats + Italian ryegrass (*Lolium multiflorum* Lam.) were
88 planted in the winters (May to October) for beef cattle grazing and remained
89 thus until now.

90 - The grazing plots have variable sizes ranging from 0.9 to 3.6 ha. Continuous
91 grazing was performed by three animals (of 200 kg initial mean live weight) per
92 plot, with regulating animals being used to maintain the grass height at 10 cm
93 (intense grazing), 20 cm (moderate grazing), 30 cm (moderate grazing), and 40
94 cm (light grazing). Plots without grazing were used as controls. The treatments
95 were arranged in a completely randomized block design with three replicates.

96 - The fertilization, which followed the recommended norms and was based on
97 the soil analysis values, consisted of grassland nitrogen fertilization and
98 soybean phosphorus and potassium fertilization at doses for yields ranging from
99 4.0 and 7.0 t/ha of grassland dry matter and 4.0 t/ha of soybeans (CQFS
100 RS/SC, 2004) (Assmann et al. 2015; Schuster et al. 2016).

2.2. Samplings and analyses

- Epiedaphic fauna was collected in four samplings: two conducted immediately after the cattle left the grassland (early November 2014 and 2015) and two immediately after the soybean harvest (early May 2015 and 2016). Nine pitfall traps were installed in each plot, totaling 27 per treatment. These traps were placed in the center of the plots located in the sampling grid. The grid consisted of three 30-m-long transects established 18 m apart from each other. Each transect had three collection points, totaling nine sample points per plot. The traps were treated with a 70% ethanol solution (v/v) and remained in the field for a 7-day period (11/05-11/11/2014; 03/29-04/6/2015; 11/1-11/7/2015; and 05/06-05/13/2016). After this period, the traps were transported to the laboratory for organism counts and identification to the Order level. The study was performed with the Authorization for Activities with Scientific Purposes (Autorização para Atividades com Finalidade Científica) number 4345-6 (Biodiversity Information and Authorization System (Sistema de Autorização e Informação em Biodiversidade - SISBIO)), issued by the Brazilian Ministry of the Environment (Ministério do Meio Ambiente do Brasil).

- In November 2014 and April 2015, soil samples were collected from the 0- to 10-cm layer in the same nine collection points of each plot, as described above, for physical and chemical analysis. The clay content (densimeter), total porosity, macroporosity, and microporosity according to a tension table (Embrapa 1997); soil density, using the volumetric ring method (Blake and Hartge 1986); and gravimetric moisture were measured to assess the soil physical traits. Total organic carbon (Walkley–Black); pH (water 1:1); P and K (Mehlich-1); Ca, Mg

125 and Al (KCl 1 mol L⁻¹); V (% base saturation); and H + Al (Toledo et al.
126 2012) were measured to assess the soil chemical traits (Table 1).

127 - Litter and plant shoot samples were collected from three points of the
128 sampling grid. The plant material was dried in a conventional oven at 65°C to
129 constant mass. Meteorological data on rainfall and air temperature were
130 gathered for 2 months in each sampling period in a meteorological station
131 installed in the experimental area (Figure 1).

132 2.3. Statistical analysis

133 - The abundance, richness, and diversity (Shannon) and the equitability (Pielou)
134 indices of the epiedaphic organisms (Odum and Barrett 2007) were calculated
135 using PAST software, version 2.17 (Palaeontological Statistics). Analysis of
136 variance (ANOVA) was performed to compare the means of abundance and
137 richness data, and the Duncan test (5% probability) was used when significant
138 differences existed, using SASM-agri software (version 3.2.4). The frequency of
139 each group was calculated in the different treatments. The organisms were
140 grouped as "Others" when the means reached values of less than 1% in all
141 treatments. Redundancy Analysis (RDA) was performed using the statistical
142 software CANOCO 4.5 (ter Braak and Smilauer, 1998) to assess the effect of
143 significant environmental (soil chemical, physical, and cover) variables on the
144 epiedaphic community in the period after grazing and in one sampling period
145 after the soybean harvest.

146 3. RESULTS

147 - A total of 65,672 individuals, divided into 20 taxonomic groups, were quantified
148 from both samplings performed after grazing. The total number of individuals
149 collected in both samplings performed after the soybean harvest was 54,274,

150 belonging to 24 groups. Thus, the results showed greater taxon richness in the
151 collections performed after the soybean harvest and a higher average number
152 of individuals per treatment in the sample collections performed after grazing.

153 - Although this is a subtropical climate region (Cfa), with marked climatic
154 variation between the four seasons, the conditions observed in the four sample
155 collections were relatively stable (Figure 1). The average air temperature was
156 only 3°C higher in collections performed after grazing than in collections
157 performed after the soybean harvest. This condition was repeated in both
158 sampling years. The rainfall was more variable, although the soil moisture
159 showed no significant changes between the four collections (averaging 27.5,
160 26.9, 25.4, and 24.5% in the four collections; Table 1).

161 - The cattle grazing intensity affected the abundance of the epiedaphic
162 organisms (Table 2). The collections performed after grazing showed that the
163 higher the grazing intensity is, the lower the abundance of these organisms will
164 be. In intense grazing, the abundance was approximately 40% lower than that
165 of the treatment without grazing, which resulted in a significant difference. No
166 significant differences occurred between the other treatments and the control,
167 despite 20, 30, and 10% mean abundance reductions in treatments with the 20-
168 , 30-, and 40-cm grazing heights, respectively. In the collections performed after
169 the soybean harvest (Table 2), the number of organisms quantified in the
170 treatment without grazing was 30% higher than the mean of the treatments with
171 grazing, albeit without significant differences between treatments, given the
172 wide variation in abundance between the years.

173 - Damage to the soil invertebrate diversity from the increased grazing intensity
174 was also shown by the decreased group richness (Table 2). After the grazing

175 period, an increased richness was observed in light grazing, without significant
176 differences from the treatments with moderate grazing and without grazing
177 (Table 2). Intense grazing resulted in the lowest richness and significantly
178 differed from light grazing. In collections performed after the soybean harvest,
179 the highest richness was observed in the treatments without grazing and with
180 light grazing (Table 2). Conversely, the treatment with the lowest diversity of
181 epiedaphic organisms was grazing to a grass height of 20 cm, which showed no
182 significant difference from the treatments at the 10- and 30-cm grass heights.

183 - The diversity indices indicate the same trend observed in abundance and
184 richness (Table 2). The assessments performed after grazing showed an
185 increased diversity of soil organisms in treatments without grazing and with
186 grazing to a grass height of 30 and 40 cm. Damage to the soil ecological
187 balance occurred in the treatments with grazing performed at higher intensities
188 (10 and 20 cm), which resulted in lower diversity indices. The collections
189 performed after soybean cultivation also showed an increased biodiversity of
190 organisms in the treatment without grazing and decreased diversity in the
191 grazing treatments.

192 - Acari, Araneae, Coleoptera, Collembola, Diptera, Hymenoptera, Hemiptera,
193 Orthoptera, Dermaptera, and Larva (unidentified larvae) were the epiedaphic
194 taxonomic groups most commonly found in the experimental area. These
195 groups were mostly sensitive to changes caused by grazing intensification and
196 showed changes in frequency with the grazing intensification. The least
197 frequent organisms were grouped into "Others" and belonged to the Annelida,
198 Chilopoda, Diplopoda, Isoptera, Blattodea, Thysanoptera, Lepidoptera,
199 Neuroptera, Mollusca, Opilione, Scorpiones, Nematoda, and Odonata taxa.

200 - The mesofaunal members were the groups most commonly found in both
201 assessment periods, and Collembola and Acari accounted for more than 60% of
202 the organisms collected in nearly all the treatments (Figure 2). In the collection
203 performed after grazing, Collembola maintained nearly the same frequency in
204 all treatments, whereas the mites had a higher frequency in grazing to a height
205 of 20 cm. However, in the collection performed after the soybean harvest, the
206 frequency of mites tended to increase, and the frequency of Collembola tended
207 to decrease with the grazing intensity.

208 - Coleoptera were the most commonly found epiedaphic macrofaunal
209 organisms (Figure 2). The frequency of this group decreased at the lowest
210 grass heights, in collections performed after both grazing and the soybean
211 harvest. On average, the frequency of Coleoptera was 18% in the control and
212 the lower grazing intensities and 10% with grazing to grass heights of 10 and 20
213 cm. Only the Hymenoptera group showed an increase in frequency with the
214 grazing intensity in the collection performed after cattle grazing. This is most
215 likely associated with a more disturbed habitat because these organisms may
216 indicate environmental imbalances. Similar to the Coleoptera group, the
217 frequency of organisms of the Araneae, Orthoptera, and Dermaptera orders
218 decreased with grazing to the lowest grass heights, thus indicating that grazing
219 intensification affects several groups of edaphic fauna.

220 - The comparison between collection periods shows that soybean cropping
221 decreased the macrofauna frequency and increased the mesofauna frequency,
222 mainly due to the increase in mites (Figure 2). On average, macroorganisms
223 accounted for 36% of the organisms collected after cattle grazing, whereas this
224 percentage decreased to 23% in the collection performed after the soybean

225 harvest. Various non-edaphic, adult individuals of the Diptera order were
226 captured in the four collections. However, their presence was associated with
227 attraction to the trap ethanol solution. Therefore, these data were disregarded.

228 - The matrix of the soil chemical and cover variables explained 22% of the
229 variation in the epiedaphic fauna data ($p=0.002$), whereas the matrix of soil
230 physical variables explained 10% of this variation ($p=0.01$) in the redundancy
231 analysis of the collection performed after grazing. The exclusive percentage
232 effect of each significant variable was 10% for the shoot dry matter, 6% for the
233 pH, 3% for the calcium, 6% for the moisture, and 3% for the soil density.
234 Conversely, the matrix of the soil chemical and cover variables only explained
235 5% of the variation in the data for the epiedaphic fauna ($p=0.05$), and the matrix
236 of the soil physical variables explained 36% of this variation ($p=0.001$) in the
237 collection performed after the soybean harvest. In this case, an exclusive
238 percentage effect of each significant variable was 5% for pH, 31% for moisture,
239 and 6% for microporosity.

240 **4. DISCUSSION**

241 - Grazing intensification for 15 years significantly decreased the epiedaphic
242 faunal diversity. In general, the control (without grazing) and the low-intensity
243 grazing treatments had the highest abundance and richness and the best
244 diversity indices. Low soil invertebrate diversity was associated with high-
245 intensity grazing treatments. Grasslands managed sustainably have
246 microhabitats with improved edaphic conditions, increased plant biomass cover
247 (Gill 2007), improved physical structure (Conte et al. 2007), increased residue
248 input (Bayer et al. 2009), and increased soil organic matter (Rosenzweig et al.
249 2016). Furthermore, the vegetation height in the grazing areas and the

250 invertebrate abundance are positively correlated ($p=0.009$) because a more
251 structured and complex vegetation cover results in better food and habitat
252 conditions (Hoste-Danyłow et al. 2010). Conversely, grazing managed
253 intensively may lead to a decreased abundance in organisms from key edaphic
254 arthropod orders due to habitat simplification (Little et al. 2013; Swengel 2001).

255 - The frequency of the edaphic mesofaunal organisms (Collembola and Acari)
256 increased in both assessment periods. These organisms are the most abundant
257 and widespread arthropods in most soils (Bedano et al. 2011). Their activity
258 improves soil quality and plant yield because they are highly active in soil
259 organic matter decomposition, nutrient cycling, and biological pest control
260 (Moreira et al. 2010). In both sampling periods, the mites were frequent at the
261 20-cm grass height. This result may indicate edaphic mite preference for areas
262 with high bovine manure availability and good soil plant cover. These conditions
263 are found at the 20-cm grass height. The plots with the lowest grass heights are
264 also the smallest and therefore have a higher input of manure because the
265 number of animals is the same in all plots. According to Silva et al. (2014), in
266 this same experimental protocol, the manure dry mass production in a grazing
267 season was 669, 478, 366, and 213 kg ha⁻¹ for the 10-, 20-, 30-, and 40-cm
268 grass height treatments, respectively.

269 - Collembola was present at a higher frequency than mites were in the period
270 after grazing in all treatments. This relation indicates a slightly disturbed
271 environment (Mateos 1992). However, the collection performed after the
272 soybean harvest showed that the plots with an increased frequency of mites
273 also showed a decreased frequency of Collembola, which also has been
274 reported in other studies (Rieff et al. 2014; Rieff et al. 2016). This behavior may

275 be associated with predation because some groups of mites are microarthropod
276 predators (Mejía-Recamier et al. 2013).

277 - The Coleoptera order is abundant in most Brazilian soils and is extremely
278 important for the soil chemical and physical quality (Correia and Oliveira 2005;
279 Pompeo et al. 2016; Portilho et al. 2011). The results from the present study
280 show the importance of an adequate grassland management for ICL
281 sustainability. Coleoptera were more commonly found at low grazing intensities
282 and in sample collections performed after grazing, regardless of the sampling
283 period. All results show that most living or dead plant cover provides a more
284 suitable habitat for the survival of these organisms (Pompeo et al. 2016; Rosa
285 et al. 2015). These effects may be direct, through increased food and shelter
286 availability, or indirect, through improved chemical and physical conditions, as
287 outlined in Table 1.

288 - Organisms of the Araneae, Dermaptera, and Orthoptera Orders were also
289 adversely affected by the increase in grazing intensity. Spiders are the most
290 diverse and abundant arthropod predators in grazing ecosystems (Horváth et al.
291 2009). Therefore, they have a key role in community structure and natural pest
292 control (Sunderland and Samu 2000). The abundance and diversity of these
293 organisms are directly associated with the vegetation vertical structure and
294 height (Bell et al. 2001; Harris et al. 2003; Horváth et al. 2009). Some
295 Dermaptera groups are important for crop production because they contribute to
296 biological pest control (Buzzi 2013). Conversely, Orthoptera are important in
297 grazing areas because they account for most of the biomass of the edaphic
298 invertebrates in those systems (Little et al. 2013).

299 - Hymenoptera (ants) are one of the bioindicators of disturbed environments
300 most studied in the literature (Pereira et al. 2007; Rocha et al. 2015). A study
301 conducted in the Brazilian Midwest showed that ICL may favor the ant
302 community, depending on the management adopted, and they may be used as
303 bioindicators in this type of soil management (Crepaldi et al. 2014). In the
304 collection performed after grazing, the frequency of these organisms was
305 highest in the treatment with 10 cm of grass height, which is the treatment with
306 the most disturbed edaphic environment.

307 - Redundancy analysis indicated that soil moisture was a key factor in the
308 composition of the community of epiedaphic organisms (Figure 3). Soil moisture
309 was higher in treatments without grazing and with moderate and light grazing in
310 both times. This performance may result from the improved edaphic conditions
311 observed in these treatments due to increased soil compaction and vegetation
312 and litter mass in the area. Intense shoot biomass extraction, as observed in
313 intense grazing, converts the soil into an environment less protected from the
314 high levels of solar radiation and therefore with increased evaporation (Gong et
315 al. 2014).

316 - Increased soil density and decreased porosity, as observed in intense grazing,
317 limit the activity of these gallery-builder organisms and the abundance of those
318 requiring porous soil space to survive (Moço 2006). Furthermore, the area fails
319 to provide quality habitats, resources, and shelter to support the high diversity of
320 edaphic organisms, given its decreased soil cover (Bedano et al. 2016). The
321 damage caused by 15 years of intense grazing (10 cm) to the soil physical
322 quality severely affected the epiedaphic fauna. Such losses were not even

323 offset by the bovine manure input per area in intense grazing, which was three
324 times higher (669 kg ha⁻¹) than that in light grazing (213 kg ha⁻¹).

325 - The pH and calcium content were the chemical factors that most affected the
326 epiedaphic fauna. The pH adversely affected the epiedaphic fauna in both
327 sample collection times. This characteristic is important to establish specific
328 groups, including ants (Jacquemin et al. 2012), and is usually inversely
329 correlated with the soil abundance of the organisms (Harada and Bandeira
330 1994; Nowak 2001). In the sample collection performed after grazing, the
331 calcium content positively affected the distribution of the edaphic organisms,
332 which may be attributed to the importance of this nutrient for the physiology of
333 some groups of invertebrates because calcium is associated with mechanisms
334 of osmotic regulation and ecdysis (Rosa et al. 2015). Soil physical, chemical,
335 and cover attributes, changed by the different grazing intensities, significantly
336 affected the epiedaphic organisms, which was also observed in other studies
337 assessing different management systems (Bartz et al. 2014; Souza et al. 2016)

338 **5. CONCLUSIONS**

339 - Grazing intensification for 15 years reduces the epiedaphic faunal abundance,
340 richness, and diversity in an integrated soybean-beef cattle production system.
341 - The Collembola, Acari, Coleoptera, Araneae, Orthoptera, Dermaptera, and
342 Hymenoptera groups are the most sensitive to the different grazing intensities.
343 - Plant cover and soil moisture are the factors that most affect the epiedaphic
344 organisms in an integrated soybean-beef cattle production system.

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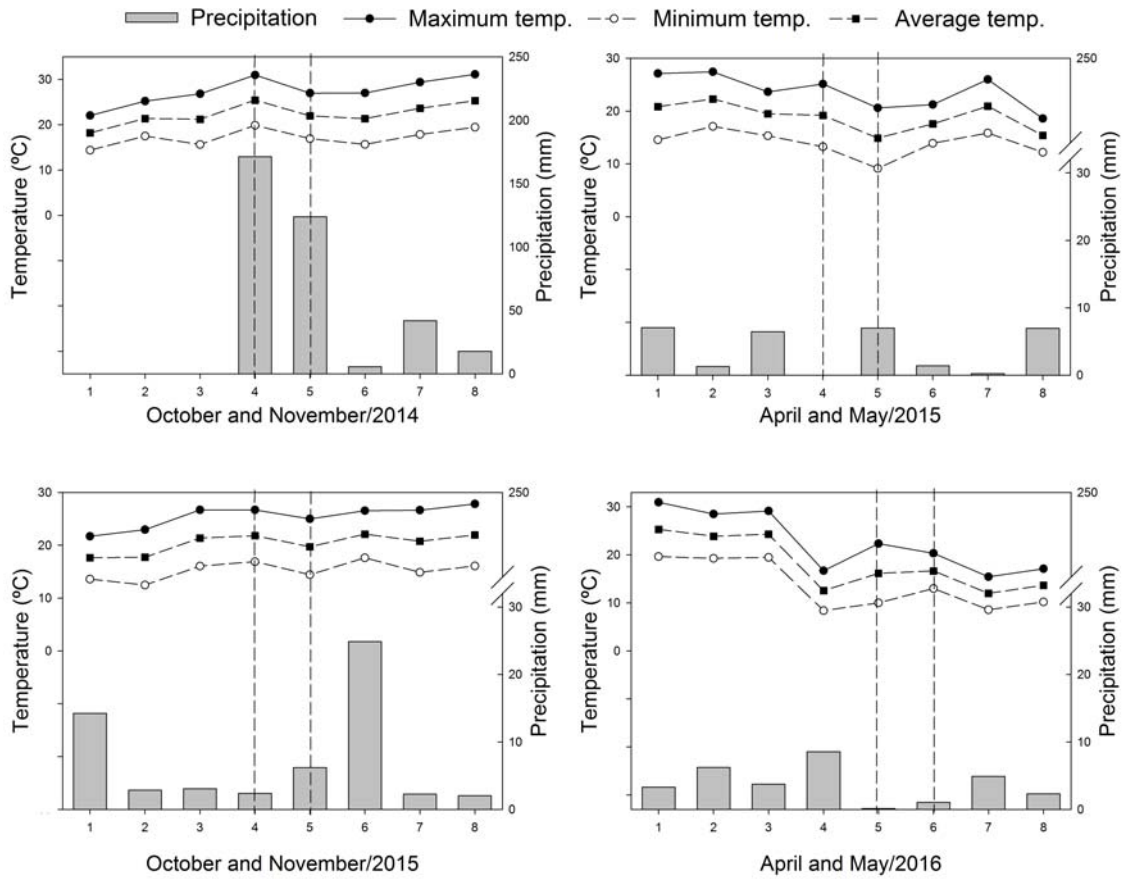
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524 Figure 1: Precipitation and minimum, mean, and maximum air temperatures in
 525 the collection periods. The period between dotted lines indicates the days on
 526 which the traps remained in the field.

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535 Table 1: Physical and chemical attributes of the 0- to 10-cm soil and plant cover
 536 layer in the treatments with different grass heights or without grazing (WG), in
 537 collections performed after cattle grazing and soybean harvest, in a 15-year
 538 integrated soybean-beef cattle production system. The data are expressed as
 539 the means of five replicates per treatment.

Treatment	Physical variables			Chemical variables										Plant cover variables				
	SD ¹ g cm ⁻³	Ma ² cm ³	Mi ³ cm ⁻³	Moisture %		pH	P mg dm ⁻³	K mg dm ⁻³	TOC ⁴ g kg ⁻¹	V ⁵ %	Al	Ca cmol _c dm ⁻³	Mg cmol _c dm ⁻³	H+Al cmol _c dm ⁻³	LDM ⁶ Mg ha ⁻¹		SDM ⁷ Mg ha ⁻¹	
				2014	2015										2014	2015	2014	2015
WG	1.26	0.13	0.42	28.9	28.9	4.3	13.1	219.3	- ⁸	27.1	1.4	3	1.2	13.6	6.6	5.1	4.6	7.3
10 cm	1.40	0.09	0.40	25.9	23.7	4.8	9.0	187.4	-	48.5	0.8	4.1	1.7	6.9	1.3	1.0	1.8	0.9
20 cm	1.37	0.09	0.41	27.4	24.6	4.9	10.7	200.4	-	50.6	0.6	4	1.8	6.4	3.0	1.8	3.3	2.6
30 cm	1.33	0.12	0.42	28.0	24.6	5.0	9.2	163.4	-	52.5	0.6	4.5	1.9	6.4	3.1	2.1	3.4	3.3
40 cm	1.32	0.14	0.38	27.3	25.4	4.8	13.6	200.6	-	48.6	0.5	4.4	1.8	7.2	3.5	4.0	4.8	4.1
				2015	2016										2015	2016	2015	2016
WG	1.34	0.09	0.45	28.4	26.4	4.3	19.5	222.6	23.1	36.4	1.3	3.9	1.5	11.2	9.2	8.5	0 ⁹	0
10 cm	1.37	0.09	0.41	25.5	23.9	4.6	9.7	136.3	22.2	44.6	1.1	4.3	2.0	8.5	4.1	4.3	0	0
20 cm	1.36	0.11	0.41	26.6	24.4	4.7	13.6	181.2	23.3	47.5	1.0	4.7	2.2	8.2	6.3	6.0	0	0
30 cm	1.36	0.09	0.43	27.0	23.9	4.8	9.4	148.3	22.8	49.0	0.7	4.6	2.1	7.5	6.6	6.0	0	0
40 cm	1.33	0.10	0.43	27.0	24.0	4.6	12.6	173.8	22.3	45.7	0.8	4.7	1.9	8.4	6.9	8.0	0	0

540 ¹ Soil density; ² Macroporosity; ³ Microporosity; ⁴ Total organic carbon; ⁵ Base saturation;
 541 ⁶ Litter dry mass; ⁷ Shoot dry matter; ⁸ Data not shown; ⁹ No growing plants remained
 542 after the soybean harvest.

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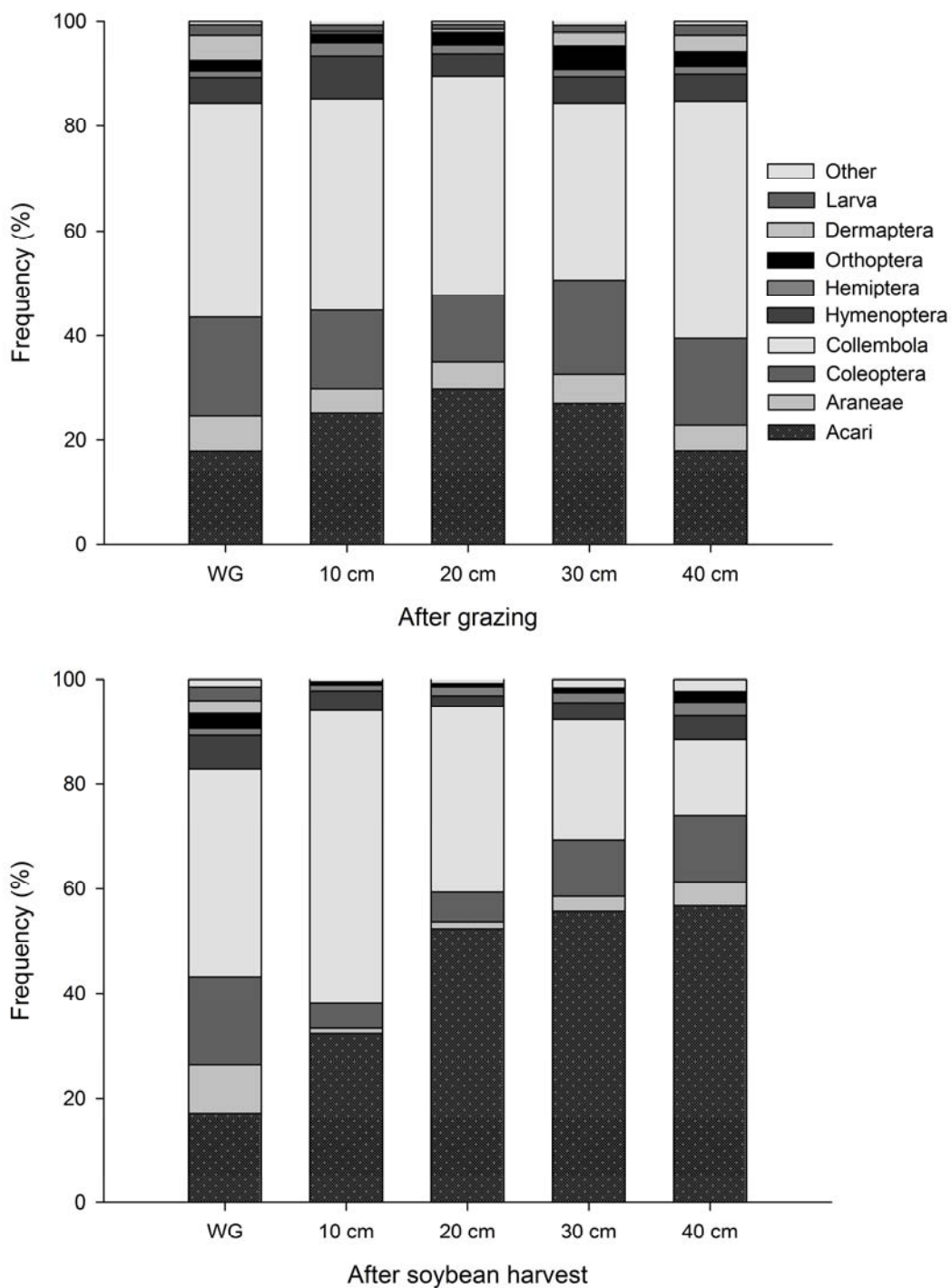
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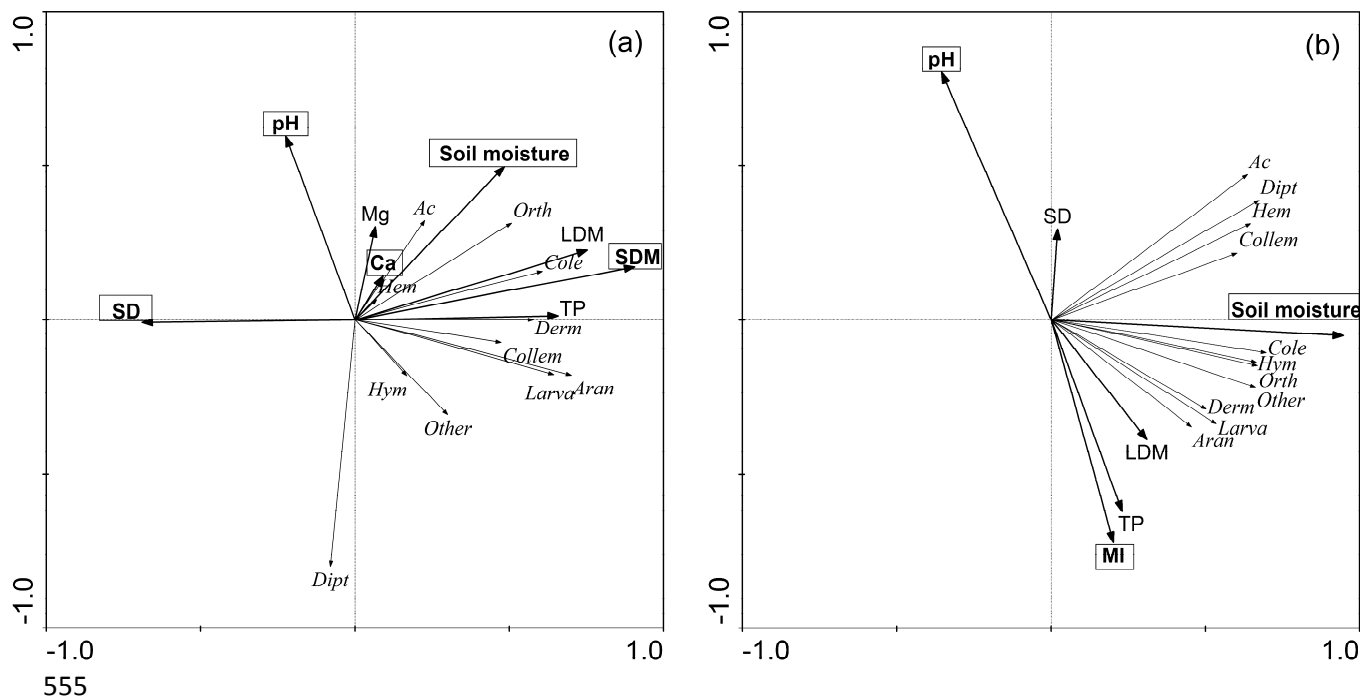
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550 Figure 2: Relative frequency of the epiedaphic fauna groups after cattle grazing
 551 (November 2014 and 2015) and after the soybean harvest (May 2015 and
 552 2016) in the treatments with different grass heights or without grazing (WG) in a
 553 15-year integrated soybean-beef cattle production system. The data are
 554 expressed as the means of 27 replicates per treatment.



556 Figure 3: Redundancy analysis (RDA) after cattle grazing (November 2014 and
 557 2015; a) and after the soybean harvest (May 2015 and 2016; b) in a 15-year
 558 integrated soybean-beef cattle production system managed with grass heights
 559 of 10, 20, 30, and 40 cm or without grazing. The variables with significant
 560 responses are highlighted with a box. Ac (Acari); Dipt (Diptera); Hem
 561 (Hemiptera); Collem (Collembola); Cole (Coleoptera); Hym (Hymenoptera); Orth
 562 (Orthoptera); Derm (Dermaptera); Aran (Araneae); Larva (unidentified larvae);
 563 Others (sum of the less frequent groups); pH; SD (soil density); TP (total soil
 564 porosity); MI (soil microporosity); TOC (total organic carbon); Ca (calcium); Mg
 565 (magnesium); LDM (litter dry mass); and SDM (shoot dry mass). The data refer
 566 to 27 replicates per treatment.

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568 Table 2: Abundance, richness, and diversity indices of epiedaphic fauna
 569 collected after cattle grazing and soybean harvest, in the treatments with
 570 different grass heights or without grazing (WG), in a 15-year integrated
 571 soybean-beef cattle production system. The data are expressed as the means
 572 of 27 replicates per treatment.

Variable/ Index	Treatments				
	WG	10 cm	20 cm	30 cm	40 cm
After grazing					
Abundance	361a	224b	283ab	256ab	324ab
Richness	12ab	11b	12ab	12ab	13a
Shannon	1.83	1.69	1.69	1.85	1.82
Pielou	0.73	0.69	0.67	0.74	0.71
After soybean					
Abundance	173ns	156	146	93	89
Richness	12a	10ab	9b	10ab	12a
Shannon	1.67	1.43	1.38	1.42	1.38
Pielou	0.68	0.62	0.63	0.62	0.56

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5 DISCUSSÃO GERAL

Os dados obtidos neste estudo resultaram de quatro amostragens realizadas ao longo de dois anos, em um experimento manejado com sistema integrado de produção agropecuária conduzido há 15 anos sob diferentes pressões de pastejo, em uma área de 23 hectares. Este trabalho é suportado por um conjunto muito grande de dados de solo, planta, animal e meteorológicos coletados ininterruptamente neste período. Constitui-se, portanto, numa excelente oportunidade para o estudo da biota do solo neste tipo de sistema, visando reduzir a grande carência de informações sobre o tema a nível mundial.

O modo produtivo atualmente utilizado por grande parte dos agricultores no sul do Brasil é um sistema puramente agrícola. No inverno, 86% área agrícola é cultivada somente com plantas de cobertura para a produção de palha para o plantio direto da soja no verão (CONAB, 2011). Este modo produtivo concentra na soja toda a renda do agricultor, o que é um grande risco, tendo em vista as instabilidades meteorológicas e mercadológicas. Por outro lado, as pastagens de invernos de aveia e azevém são altamente produtivas e podem ser utilizadas para o pastejo animal e ainda produzirem biomassa vegetal suficiente para o plantio direto do verão (KICHEL e MIRANDA, 2006; VILELA et al., 2008).

Frente a este contexto, a introdução do pastejo bovino no inverno representa a busca por aumento e estabilidade de renda aos agricultores, com intensificação uso do solo (BERRY et al., 2003). Porém frequentemente questiona-se se a introdução do pastejo pode prejudicar o solo. Para os organismos do solo, o pastejo pode aumentar sua diversidade, atividade e biomassa, pois há introdução de maior quantidade e diversidade de resíduos através da urina e esterco incorporados pelos animais (CLEGG, 2006; LIN et al., 2009). Além disso, a atividade e o crescimento radicular das forrageiras são incentivados pela desfolhação provocada pelo pastejo, o que gera um incremento na taxa de rotatividade de rizodeposições e decomposição de raízes finas (HEWINS et al., 2016; PAPATHEODOROU et al., 2008). Por outro lado, o pastejo pode prejudicar a atividade biológica, pois promove compactação da camada superficial do solo, reduz a porosidade e por consequência a aeração e a infiltração de água, retira parte da biomassa vegetal, o que altera a temperatura e a umidade do solo (JIA et al., 2006; SOUTO et al., 2008).

Preocupados com a sustentabilidade dos SIPAs, buscou-se neste trabalho conhecer as implicações das diferentes alturas de pastejo dos bovinos sobre a

comunidade de organismos edáficos. Foram avaliadas a atividade e a biomassa microbiana através da taxa de respiração basal do solo e do carbono da biomassa microbiana, e a diversidade de invertebrados epiedáficos. Estes parâmetros foram escolhidos por serem bioindicadores de alta sensibilidade da qualidade do ambiente e porque uma alta atividade biológica do solo é requisito para a sustentabilidade de qualquer sistema produtivo (BARETTA et al., 2014; CLUZEAU et al., 2012; PONGE et al., 2013). Nos solos biologicamente ativos, os processos de degradação dos resíduos orgânicos, mineralização dos nutrientes, formação da matéria orgânica do solo, agregação do solo, controle biológico, produção de fitohormônios, etc., ocorrem em uma magnitude satisfatória, contribuindo para a aumentar a produtividade e preservar o ambiente (BARTZ et al., 2014; SHARKHUU et al., 2016).

Para este trabalho, tanto os parâmetros microbiológicos quanto a diversidade da comunidade epiedáfica responderam aos tratamentos aplicados. Dos atributos do solo, a umidade foi determinante para a composição e atividade biológica no solo, assim como a cobertura vegetal. Quanto maior a pressão de pastejo aplicada, menor foi a cobertura de biomassa vegetal, a serapilheira e a umidade do solo. Nestes tratamentos de pastejo mais intenso foram observadas a menor taxa de respiração basal e o menor teor de carbono na biomassa da comunidade microbiana, assim como uma menor diversidade da comunidade epiedáfica. O pastejo manejado de forma intensiva pode levar a um decréscimo da abundância e atividade da biota do solo devido a simplificação de hábitat (LITTLE et al., 2013; PERI et al., 2015; SWENGEL, 2001), o que pode acarretar em prejuízos importantes para a qualidade do solo.

As avaliações microbiológicas mostraram que tanto a taxa de respiração basal do solo, quanto o conteúdo de carbono da biomassa responderam às intensidades de pastejo e aos distúrbios ocasionados por ela. Quanto maior foi a pressão de pastejo aplicada, menor foi a atividade microbiana. Desta forma, estes atributos podem ser utilizados como bioindicadores de qualidade do solo, assim como observado em outros estudos avaliando diversos indicadores microbiológicos em outros sistemas agrícolas (FERREIRA et al., 2011; LISBOA et al., 2012; MOESKOPS et al., 2012; MOSCATELLI et al., 2012; VASCONCELLOS et al., 2013). Da mesma forma alguns grupos integrantes da fauna epiedáfica do solo também podem ser utilizados como bioindicadores da qualidade do SIPA (BARETTA et al., 2014). Os ácaros, colêmbolos, coleópteros, aranhas e ortópteras se mostraram sensíveis às pressões de pastejo aplicadas, sendo assim, adequados ao uso como bioindicadores neste sistema.

Nas maiores alturas de pastagem os prejuízos causados pelo pastejo bovino aos organismos do solo são minimizados. A maior biomassa vegetal viva ou morta nestes tratamentos proporciona melhores condições físicas do solo, que resulta em maior umidade do solo e por consequência maior atividade, biomassa e biodiversidade dos organismos. Estes resultados reforçam a necessidade de um ajuste adequado da carga animal na pastagem para que não ocorra comprometimento à sustentabilidade dos SIPAs.

6 CONCLUSÃO GERAL

A partir dos resultados obtidos nos dois estudos conduzidos em um SIPA de longa duração pode-se concluir que:

Há maior respiração microbiana e maior conteúdo de carbono na biomassa microbiana do solo no tratamento sem pastejo e quando o pastejo ocorre nas maiores alturas da pastagem;

Na menor altura da pastagem há prejuízos aos atributos físicos do solo, o que resulta em menor teor de umidade, e redução da atividade e da biomassa dos microrganismos;

A intensificação do pastejo reduz a abundância e a diversidade da fauna epiedáfica;

A perda da qualidade física do solo sob pastejo intenso resulta em menor teor de umidade e este fator é o que mais contribui para a redução da abundância e da riqueza dos meso e macrorganismos;

Os atributos biológicos do solo avaliados no presente estudo podem ser utilizados como bioindicadores de qualidade do solo em um SIPA de longa duração.

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